Characteristics and sources of halogenated hydrocarbons in the Yellow River Delta region, northern China

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ABSTRACT

As important ozone-depleting substances and potential greenhouse gases, halocarbons have attracted much attention since 1970s. To better understand the characteristics and sources of halocarbons in the Yellow River Delta (YelRD) region, an important oil extraction and petrochemical industrial area in northern China, a two-phase (winter-spring and summer) intensive campaign was conducted in 2017. Compared with the Northern Hemispheric background, median enhancements of Montreal Protocol (MP) regulated species (CFCs, Halons and CH₃CCl₃) in this region were relatively small. Meanwhile, enhancements were relatively large for MP under control species (HCFCs and HFC-134a). Historical trends of halocarbons in China and the YelRD region both exhibited declining and increasing trends for MP regulated and under control species, respectively. Results above indicate effective implementation of MP in China. The enhancements of MP unregulated species were high as they are still used and uncontrolled. In addition, oil fields were found to have a negligible influence on ambient halocarbons in this region. The back-trajectory analysis shows that clusters from the North China Plain and East China were associated with the highest concentrations of most halocarbons during the winter-spring and summer periods, respectively. Source apportionment of halocarbons was carried out by positive matrix factorization (PMF). Model results indicate that biomass burning, marine, refrigerants and foam blowing, and solvent usage were the most important sources. This study provides the underlying insights into the characteristics and sources of halocarbons in the YelRD region.

1. Introduction

Since Molina and Rowland (1974) revealed the mechanism of chlorine atom catalysed destruction of stratospheric ozone in 1970s, halocarbons have got extensive attention as ozone depleting and global warming substances (IPCC/TEAP, 2005). Halocarbons can be divided into several sub-classes, including chlorofluorocarbons (CFCs), Halons, hydrochlorofluorocarbons (HCFCs), hydrofluorocarbons (HFCs), and some short-lived halogenated substances (a broad class of chlorine- and bromine-containing substances known as very short-lived halogenated substances (VSLs) that have lifetimes shorter than about 6 months and are not controlled under the Montreal Protocol). CFCs, along with CH₂Cl₂ and CCl₄ were widely used in the pre-MP era (prior to the signing of the Montreal Protocol) as refrigerants, closed-cell foam plastics, aerosol propellant, medical aerosol products, solvents, and industrial applications (Artuso et al., 2010; McCulloch and Midgley, 2001; McCulloch et al., 2003; Sturrock, 2002). Due to their long lifetimes, CFCs can be transported to the stratosphere where they could be photolyzed to release the highly reactive chlorine/bromine atoms and destroy ozone (WMO, 2014). Halons were also widely applied in the pre-MP era as fire extinguish agents in domestic, commercial and military applications (Montzka et al., 1999). However, the ozone depletion potentials (ODPs) of Halons are significantly larger than that of CFCs because bromine is much more effective (about 60 times) in
destroying ozone (WMO, 2014). In the past few decades, HCFCs were increasingly used in refrigeration, air conditioning and foam plastics to replace CFCs for their lower ODPs (Kim et al., 2011). Meanwhile, HFCs have nearly zero ODPs and thus are considered as better substitutes for CFCs and HCFCs (O’Doherty et al., 2004). Despite of their differences in ODPs, all the aforementioned halocarbons are greenhouse gases and were estimated to contribute positive globally averaged radiative forcing (IPCC/TEAP, 2005). Moreover, most of the VSLs are toxic to human (Simmonds et al., 2006) and some VSLs may also contribute significantly to ozone loss in the stratosphere (Fang et al., 2018). After realizing the environmental effects of halocarbons, a series of treaties, including Montreal Protocol on Substances that Deplete the Ozone Layer and its amendments, were contracted to control the ODSs. After years of control, the overall abundances of ODSs in the global atmosphere have kept decreasing for more than a decade (WMO, 2014). Levels of CFCs, CCl4, CH3CCl3, and H-1211 (Halon 1211) continue declining, while H-1301 is expected to maintain a certain level due to its continued small releases and long lifetime (65 years). CH3Br concentrations have decreased by 98% compared to its peak levels (WMO, 2014). Concentrations of HCFCs are still increasing, but total emissions of major HCFCs have declined as a result of the implementation of MP. HFCs have been increasing at a relatively rapid rate because of great application demands (WMO, 2018). Nevertheless, some unexpected phenomena also occurred. Although CFC-11 mixing ratios are still declining, its emissions may have increased in recent years (Montzka et al., 2018). Lunt et al. (2018) found continued emissions of the CCl4 from eastern Asia. Unexpected emissions of CFC-113, CFC-114 and CCl3 after 2012 were reported (WMO, 2018). A recent study reported that if emissions of CHCl3 (chloroform, belonging to VSLs) continue to grow at the recent rate, future ozone-layer recovery could be delayed by several years (Fang et al., 2018a). In summary, the implementation of MP is effective, but we should still be vigilant for unexpected emissions and their environmental impacts.

China has made great progress on the phase-out of halocarbons and eliminated more than 100,000 tons of ODSs in the past 30 years (Fang et al., 2018b). With all ODSs being phased out on schedule, CFCs, Halons and CCl4 were even eliminated in advance in 2007 (Wu et al., 2018). Previous studies have reported the characteristics of halocarbons in China. Fang et al. (2012a) estimated that total emissions of CFCs and HCFCs were decreasing in China between 2001 and 2009. Two other studies showed that most of the halocarbons in 45 Chinese cities were only within 10% to 20% enhancements of the global background (Barletta et al., 2006; Fang et al., 2012b). In an aircraft measurement over east Shandong Peninsula in 2011, CFCs, together with CCl4 and CH3CCl3, showed slightly higher levels compared to the global background (Wang et al., 2014a). Zhang et al. (2017) reported that CFCs showed a decreasing trend at four regional background sites of China. However, such studies became relatively rare after 2012 in China. Since the phasing out of HCFCs in China carried out in 2013, more research about halocarbons is urgently needed.

With a population of about 10 million people, the Yellow River Delta (YelRD) region is the second largest estuary delta region in China. Adjacent to the Bohai Sea, the YelRD region is located at the junction of Shandong Peninsula and Beijing-Tianjin-Hebei metropolitan area (the biggest urbanized region in Northern China). In the past 30 years, industrialization and urbanization of the YelRD region may have resulted in high pollution levels of halocarbons (Guo et al., 2009). Endowed with abundant oil and natural gas resources, many petrochemical industries which produce high levels of volatile organic compounds are located in this region. In addition, some marine source species (e.g., CH4, CH3Br, etc.) could be abundant in the coastal sea area. Thus, a better characterization of the pollution features and sources of halocarbons in this area is needed. In this study, a two-phase intensive campaign was conducted in 2017. A total of 111 and 22 samples were collected at a field site and in oil fields (all located in the YelRD region), respectively, and 24 atmospheric halocarbon species were detected. In combination with the HYSPLIT back trajectory and positive matrix factorization (PMF), we report the characteristics and sources of halocarbons in this region. In addition, considering the fact that the published halocarbons data in China is a little lacking after 2012, we provide an overview of pollution levels of halocarbons in China. We also focused more on HCFCs, because their phase-out schedule was carried out in 2013 in China. In addition, our work can provide some implications for decision-makers.

2. Methods

2.1. Sampling site

The sampling site is in the Yellow River Delta Ecological Research Station of Coastal Wetland (37.75’S, 118.97’E, 1 m above sea level), which is located in the Yellow River Delta National Natural Reserve and is only 10 km away from the Bohai Sea (see Fig. 1). There are no densely populated and industrial areas within a distance of 10 km.
around the station. Dongying City, Kenli District, and Huanghekou Town are three major cities or towns around. They are located 32 km, 42 km and 16 km to the southwest of the site, respectively. The Shengli Oil Field, which is the third largest oil field in China, surrounds the site, leaving only a narrow gap in the southeast. To sum up, the station is a complicated coastal rural site in oil fields that may receive different types of air masses which include oilfield air, marine air, urban plumes, and aged air masses transported over a long distance from the Beijing-Tianjin-Hebei region or East Asia. A detailed description of this study site is given in Zhang et al. (2019).

2.2. Sampling and laboratory analysis

A two-phase campaign was conducted in the winter-spring (from February 13 to April 3) and summer (from June 8 to July 9) of 2017. Two-liter pre-vacuumed, electro-polished, stainless-steel canisters from the University of California, Irvine (UCI) were used to collect whole air samples. During the sampling, the bellows valve of the canister was kept slightly open for 3 min to avoid the instantaneous air mass pollution. During the first 6 days of the winter-spring campaign, 2 samples were collected each day at 00:00 and 12:00 local time (LT). After February 18, 6 samples were collected each day, but only on polluted days (according to the air quality forecast from https://map.zq12369.com/) with a 3-h interval from 06:00 to 21:00 LT. In the summer campaign, 7:00, 10:00, 12:00, 14:00, 16:00, and 19:00 LT were chosen as the sampling time on polluted days. Additional 11 air samples were collected in the oil fields during each phase of the campaign. In summary, a total of 133 samples (64 for the winter-spring period and 69 for the summer period) were collected.

After the campaign, the canisters were sent back to UCI for composition and concentration analysis. The whole air samples were analysed by GC-FID/ECD/MS. Six column-detector combinations were used, which include two electron capture detectors (ECD, sensitive to halocarbons and alkyl nitrates), two flame ionization detectors (FID, sensitive to hydrocarbons), one quadrupole mass spectrometer detector (MSD, for unambiguous compound identification and selected ion monitoring) and one nitrogen phosphorus detector (NPD, for detection of nitrogen species). A total of 24 halocarbon species (see Table 1), DMS and CO were detected and analysed in this study. The accuracy and measurement precision vary by compound and are estimated to be 3–20% and 3–10%, respectively (Colman et al., 2001; Simpson et al., 2010). The limit of detection (LOD) also varies by compound, from 0.01 ppb for chlorobrominated species to 10 ppb for CFC-12 (Simpson et al., 2010). Detailed information about laboratory analysis, detector and precision of each compound can be found on the website of the Rowland-Blake Group (https://ps.uci.edu/~rowlandblake/research_lab.html) and is also given in the Supplementary material.

2.3. Backward trajectory analysis

The Hybrid Single-Particle Lagrangian Integrated Trajectory model (HYSPLIT, version 4.9; Draxler and Hess, 1998) developed by the National Oceanic and Atmospheric Administration (NOAA) Air Resources Laboratory (http://www.arl.noaa.gov/ready/hysplit4.html) was used to understand the air masses history in this study. The meteorological data were from the Global Data Assimilation System (GDAS) data set (http://ready.arl.noaa.gov/archives.php). Considering the long lifetimes of most detected halocarbons, 120-h back trajectories and associated clusters were performed and analysed. Three hundred meters above the ground level was chosen as the end point to avoid the influence of surface topography (Stohl et al., 2003). Back trajectories were restarted every 3 h (from 0:00 to 21:00 LT) and those corresponding to the sampling time were used for cluster analysis. The cluster analysis was performed with the HYSPLIT built-in cluster function which uses multivariate statistics to group the most similar trajectories into one category (Sun et al., 2016). Finally, the numbers of clusters were selected according to the total spatial variance change in scree plot.

2.4. PMF receptor model

Positive matrix factorization (PMF5.0) was used for source apportionment of halocarbons (EPA, 2014). For this study, the model was driven by 111 site samples and 25 selected species (22 halocarbons, DMS, CO and the sum of all input species). DMS and CO were treated as marine and combustion indicators, respectively (Guo et al., 2009). H-1211 and CH3CCL were excluded from the model as they exhibited slightly lower levels than Northern Hemispheric (NH) backgrounds in the YelRD region. Possible reasons for their low values will be discussed in Section 3.1. CFC-114 was also excluded from the model because it displayed a non-normal curve in the model. Due to the long lifetimes of some halocarbons, the global and regional backgrounds have significant impacts on their concentrations. Thus, the global baseline subtraction method (subtracting global or hemispheric average values from observed concentrations) was used to eliminate this effect in this study (Guo et al., 2009; Li et al., 2014; O’Doherty et al., 2001). Prior to using the method, samples with extremely high and low values classified by Tukey’s test were replaced with the measured median values. For species (CO, DMS, CHBrCl, CHBrCl2 and C2H4Cl2) whose background concentrations are not available, the corresponding 5% percentile of site samples was used. After subtracting the NH background, negative values were replaced with half limit of detection (LOD) (Sarkar et al., 2018). The uncertainty of each value was given by (EPA, 2014):

\[
\text{Uncertainty} = \sqrt{(\text{ErrorFraction} \times \text{concentration})^2 + (0.5 \times \text{LOD})^2}
\]

For negative values, uncertainties were estimated as 5/6 of LOD (Li et al., 2014).

After data pre-treatment as described above, different numbers of sources (from 2 to 10) were tested in the PMF model to find the optimal value for most reasonable results. Finally, 5 factors were chosen for both goodness-of-fit of the model and prior knowledge about halogenated compounds emissions (Li et al., 2014). Details of the identification of major sources and their contributions to the ambient air will be discussed in Section 3.3.

3. Results and discussion

3.1. General characteristics of halocarbons in the YelRD region

Table 1 shows the mixing ratios of measured halocarbons at the sampling site and in the oil fields. NH background values, p-values (for the differences between our study and the NH background), lifetime, ODPS and 100-year GWPs of each species are also listed. The NH background data were derived from the World Data Centre for Greenhouse Gases (WDGGG). Median values of measurements during February 13 – April 3 and June 8 – June 9 of 2017 at three NH sites, i.e. Trinidad Head (41.05’N, 124.15’E), Mace Head (53.32’N, 9.90’W), and Junfruojoch (46.55’N, 7.99’W), were used as the background. These background sites are located in the uninhabited mountains or capes of the Northern Hemisphere and are at similar latitude to that of the sampling site in this study. As shown in Fig. 2, the measured halocarbons are divided into 3 categories (MP eliminated species, MP under control species, and unregulated species) according to the Montreal Protocol. The first category includes CFCs, Halons, CH2CCl4, CC4 and CH2Br. Median concentrations of H-1211 and CH3CCL in the YelRD region were comparable to the NH background, while the rest of the species in the first category showed slightly higher median concentrations (except CHBr and CCl4). Low concentrations of MP eliminated species (except CH3Br and CCl4) in the YelRD region demonstrate effective emission reductions of them in China. The enhancement
(~32%) of CCl₄ suggests its sources in this region. Significant enhancement was observed for CH₃Br with median concentrations of 31 (winter-spring) and 15 (summer) ppt, compared to the NH background of 7 ppt. Recent research showed that the natural oceanic source is now comparable to the oceanic sink (WMO, 2014). So, the abundance of CH₃Br in the YelRD region may be influenced by freshwater wetlands and coastal saltmarshes (WMO, 2014). This is also supported by the wind direction analysis which will be discussed later. Being widely used as refrigerants and foam blowing agents, HCFCs and HFC-134a (MP under control species) showed a dramatic increase in developing countries in the past years (Kim et al., 2011). Enhancements of HCFCs were moderate (ranging from approximately 41 to 70%) in this study.

**Table 1**

<table>
<thead>
<tr>
<th>Species</th>
<th>Ambient air (n = 111)</th>
<th>Oil field (n = 22)</th>
<th>NH background¹</th>
<th>P value ² between site and background</th>
<th>Lifetime³</th>
<th>ODPs</th>
<th>100-year GWPs</th>
</tr>
</thead>
<tbody>
<tr>
<td>CFC-12</td>
<td>530 (530 ± 12)</td>
<td>530 (529 ± 14)</td>
<td>535 (539 ± 20)</td>
<td>511 (510-518)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CFC-11</td>
<td>257 (271 ± 57)</td>
<td>269 (274 ± 26)</td>
<td>283 (302 ± 79)</td>
<td>228 (227-229)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CFC-113</td>
<td>77 (77 ± 5)</td>
<td>82 (83 ± 5)</td>
<td>78 (78 ± 3)</td>
<td>73 (72-73)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CFC-114</td>
<td>16.4 (16.5 ± 0.5)</td>
<td>16.8 (16.9 ± 0.6)</td>
<td>16.9 (17.1 ± 1.0)</td>
<td>16.3 (16.3-16.3)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>H-1211</td>
<td>3.6 (5.5 ± 7.7)</td>
<td>3.4 (3.5 ± 0.2)</td>
<td>3.4 (7.7 ± 18.7)</td>
<td>3.6 (3.6-3.6)</td>
<td>0.028</td>
<td></td>
<td></td>
</tr>
<tr>
<td>H-1301</td>
<td>3.5 (4.0 ± 3.7)</td>
<td>3.5 (3.6 ± 0.3)</td>
<td>3.5 (3.7 ± 0.5)</td>
<td>3.4 (3.4-3.4)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CCl₄</td>
<td>109 (129 ± 62)</td>
<td>106 (117 ± 37)</td>
<td>102 (110 ± 32)</td>
<td>81 (80-81)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CH₃Br</td>
<td>31 (68 ± 98)</td>
<td>17 (25 ± 23)</td>
<td>15 (22 ± 17)</td>
<td>7 (7-8)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CH₂CCl₃</td>
<td>2.8 (2.9 ± 0.4)</td>
<td>2.6 (2.6 ± 0.1)</td>
<td>2.6 (2.6 ± 0.2)</td>
<td>2.7 (2.6-2.4)</td>
<td>0.198</td>
<td></td>
<td></td>
</tr>
<tr>
<td>HFC-134a</td>
<td>104 (131 ± 100)</td>
<td>119 (122 ± 16)</td>
<td>120 (126 ± 28)</td>
<td>84 (81-97)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>HCFC-22</td>
<td>371 (1133 ± 2932)</td>
<td>427 (464 ± 184)</td>
<td>460 (469 ± 175)</td>
<td>229 (225-247)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>HCFC-142b</td>
<td>29 (31 ± 6)</td>
<td>33 (42 ± 36)</td>
<td>33 (36 ± 14)</td>
<td>22 (21-22)</td>
<td>0.000</td>
<td></td>
<td></td>
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<tr>
<td>HCFC-141b</td>
<td>36 (173 ± 564)</td>
<td>43 (45 ± 11)</td>
<td>46 (45 ± 12)</td>
<td>24 (24-26)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CHCl₃</td>
<td>283 (476 ± 701)</td>
<td>260 (378 ± 310)</td>
<td>212 (452 ± 477)</td>
<td>12 (5-14)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CH₂Cl₂</td>
<td>396 (500 ± 389)</td>
<td>574 (723 ± 467)</td>
<td>614 (624 ± 474)</td>
<td>60 (60-63)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>C₂HCl₃</td>
<td>20 (149 ± 411)</td>
<td>24 (97 ± 360)</td>
<td>22 (52 ± 81)</td>
<td>3 (3-3)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>C₂Cl₂</td>
<td>18 (50 ± 102)</td>
<td>21 (27 ± 20)</td>
<td>16 (30 ± 33)</td>
<td>4 (3-4)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CH₂Cl</td>
<td>1106 (1879 ± 1336)</td>
<td>1242 (1879 ± 1336)</td>
<td>1368</td>
<td>584 (573-603)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CH₃I</td>
<td>2.0 (2.1 ± 0.9)</td>
<td>3.4 (3.8 ± 2.2)</td>
<td>3.5 (3.0 ± 1.6)</td>
<td>0.7 (0.5-0.8)</td>
<td>0.023</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CH₂Br₂</td>
<td>1.1 (1.3 ± 0.5)</td>
<td>1.1 (1.4 ± 0.9)</td>
<td>0.9 (1.0 ± 0.3)</td>
<td>0.9 (0.8-0.9)</td>
<td>0.222</td>
<td></td>
<td></td>
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<tr>
<td>CHBrCl₂</td>
<td>1.6 (2.3 ± 2.5)</td>
<td>1.4 (1.8 ± 1.2)</td>
<td>1.3 (2.0 ± 1.7)</td>
<td>1.2 (1.4 ± 1.0)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CHBrCl</td>
<td>1.3 (1.9 ± 1.8)</td>
<td>1.2 (1.4 ± 1.0)</td>
<td>0.8 (1.5 ± 1.4)</td>
<td>1.5 (1.3 ± 1.5)</td>
<td>0.000</td>
<td></td>
<td></td>
</tr>
<tr>
<td>CHBr₂</td>
<td>3.9 (5.7 ± 4.4)</td>
<td>2.6 (4.8 ± 3.6)</td>
<td>2.5 (4.1 ± 4.2)</td>
<td>0.7 (0.6-0.9)</td>
<td>0.097</td>
<td></td>
<td></td>
</tr>
<tr>
<td>C₂H₄Cl₂</td>
<td>480 (732 ± 717)</td>
<td>370 (586 ± 511)</td>
<td>238 (334 ± 304)</td>
<td>0.6 (0.5-1.0)</td>
<td>0.076</td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

¹ Units are in ppt.

² Using the latest data that can be obtained for the three NH background sites (Trinidad Head (41.05°N, 124.15°E), Mace Head (53.32°N, 9.90°W), Junfraujoch (46.55°N, 7.99°E)), the NH background medians were calculated for dates February 13 – April 3 and June 8 – June 9 of 2017, which is consistent with the sampling time of this study. The year of the background data is species dependent, ranging from 2008 to 2017.

³ The atmospheric lifetime, ODPs, GWPs of halogenated hydrocarbons are from the Scientific Assessment of Ozone Depletion (WMO, 2014).

⁴ Very low values are not listed.

Fig. 2. Median enhancement of each halocarbon species compared with the NH background in the YelRD region.
due to the fact that their stepwise phase-out has been in force since 2013 in China. The rest chlorinated and brominated halocarbons are not regulated by MP and are categorized as unregulated species. Their enhancements were significant (ranging from approximately 20% to 2100%), indicating strong natural and/or anthropogenic sources in this region as their lifetime is relatively short.

Seasonality appears in some halocarbons due to differences in meteorological conditions, OH radical levels, and boundary layer height (Wang et al., 2005). Table 1 shows that mixing ratios of MP eliminated species (except CH₂Br) exhibited no seasonal differences because they possess a long lifetime and have been eliminated for 10 years. Wingenter et al. (1998) reported that the lowest concentrations of CH₂Br were observed in summer in the Northern Hemisphere. CH₂Br mixing ratios were also found to be lower in summer (the median was 13.5 ppt lower than that in winter-spring) in this study. This could be caused by the strengthening of its sinks in the summertime. First, reaction with OH is the dominant removal pathway of CH₂Br in the troposphere and the removal rate is lower in the winter-spring period (Blake et al., 1997). Second, there is more additional loss of CH₂Br in summer, which includes photolysis and soil microbial degradation (Wingenter et al., 1998; WMO, 2014). HFCs (HCFC-22 and HCFC-142b) and HFC-134a concentrations were higher in summer, which was probably due to the extensive use of air conditioning. CH₂Cl levels were higher in summer as a result of biomass burning (Xue et al., 2011), CH₂Cl₂, C₂HCl₃, and C₂Cl₄ were widely used in industry as solvents and degreaser (Sturrock, 2002) and exhibited lower concentrations in the winter-spring period. This is contrary to the previous studies which reported higher concentrations of these species in winter (Artuso et al., 2010; Wang et al., 2005; Wingenter et al., 1998). This difference may be caused by the emission control policy in China. To prevent severe haze pollution, many factories, including electronic factories, PCB factories etc., were required to reduce production in winter until next spring. Marine is the major source of CH₃I. Mixing ratios of CH₃I were found to be higher in summer because of higher sea surface temperature and the summer monsoon which brought more marine air to the sampling site (Yokouchi et al., 2011). In summary, due to the influences of chemistry, meteorological conditions and emissions, some halocarbons exhibited an obvious seasonality in the YelRD region.

Comparing data collected at the sampling site with those in the oil fields and refinery, mixing ratios of most halocarbons were comparable (Table 1). Because oil production, transport and refining processes did not use or produce halocarbons, the influence of oil fields and refinery on these species was minor. For MP eliminated species (except CH₂Br which was lower in the oil fields), concentrations at the site and in the oil fields were similar, indicating an even distribution. But owing to the fact that oil fields are adjacent to Kenli District, some industrial and terrestrial biogenic source species, such as HCFC-22, HCFC-141b, CH₂Cl₂, and CH₂Cl, exhibited higher levels in the oil field samples with enhancements ranging from 13% to 29%. Meanwhile, because the oil fields are 25 km further from the Bohai Sea than the sampling site, levels of some marine source species (CHCl₃, CH₂Br₂, CHBrCl₂, CHBr₂Cl and CHBr₃) were lower in the oil fields, ranging from 14 to 33%. Comparison of halocarbon levels in different locations reflects the fact that the oil extraction and refining should not be an important source of halocarbon compounds.

Wind sectors combined with halocarbon levels can provide insights into the local sources at a small scale. Eight species of halocarbons are presented in Fig. 3 to demonstrate the impacts of local sources. The predominant wind was from the south during the whole study period. Concentrations of CFC-12 (representative of CFCs, Fig. 3a), Halons and CH₂CCL₃ (figures not shown) were not affected by wind directions, indicating that they were from the residues in the atmosphere rather than local emissions. CH₂Cl₂, CHCl₃ and CH₂Cl (Fig. 3b-d), together with CHBrCl₂, CHBrCl₂, CHBr₂Cl, C₂H₄Cl₂ and CCl₄ (figures not shown), showed the highest values in the S-SW and S-SE sectors, suggesting the influence of the upwind urban areas to the south or (and) the Bohai Sea to the southeast. Their regional sources include marine, marine biogenic, biomass burning and industrial solvents (Chan and Chu, 2007; Simmonds et al., 1998; Wadden et al., 1991). C₂HCl₃, CCl₄ and HCFC-22, species that have been widely used in industry applications, such as metal degrease, solvents, and refrigerants (Simmonds et al., 1998; Wadden et al., 1991), exhibited peak values in the W-SW and S-SW sectors. These results suggest the impacts of regional sources, especially the ones associated with the southwest sector where Dongying City, Kenli District and Huanghekou Town are located, on many halocarbons at the study site. Peak levels of CH₃I (Fig. 3h), CH₂Br and CHBr₂Cl (figures not shown) appeared in the S-SE sector and they might be influenced by freshwater wetlands, coastal saltmarshes and the Bohai Sea (WMO, 2014).

Fig. 4 demonstrates the diurnal variations of halocarbons in the two sampling periods. As presented in Fig. 4a, the MP eliminated species (except CCL₃) showed negligible variations because they have been phased out for 10 years in China. CCl₄ showed slight diurnal variations and exhibited its maximum (141.1 ppt) and minimum (116.3 ppt) at 12:00 and 18:00 LT, respectively. This suggests the existence of CCl₄ sources in this region. In Fig. 4b, the MP under control species showed higher levels in nighttime or early morning. Meanwhile, variations of these species were negligible from 9:00 to 18:00 LT. This is because these species have regional sources and would accumulate at night due to the lower boundary layer height (Guo et al., 2009). Fig. 4c exhibits the diurnal variations of CH₂Br and some unregulated species. Mixing ratios of these species showed the maximum and minimum in the morning and afternoon, respectively. This should be associated with their strong local sources and the evolution of the boundary layer height.

China has experienced dazzling economic booming during the past decades and the implementation of Montreal Protocol has been effective at the same time. Thus, various halocarbons exhibited quite different historical varieties for both reasons. Fig. 5 shows the varieties of some halocarbons in China, with different regions plotted in various colours. Up to now, CFC-11, CFC-12, CFC-113, CCL₃, CH₂CCL₃ showed a decline of 4.2, 6.6, 1.2, 1.3 and 4.2 ppt per year since 2000, respectively. The corresponding decline rates of the NH background were 1.8, 1.7, 0.7, 1.1 and 2.5 ppt per year. The production of these species should have been totally banned for about 10 years in China and their decline rates were 1.2 to 3.9 times faster than the NH background, indicating effective implementation of MP by the Chinese Government. CFC-114 is restrictively applied as propellant and refrigerant but rarely used in China with an estimated emission of 0.1 ± 0.02 Gg/yr for 2008–2009 in the PRD region (Wu et al., 2018; Zhang et al., 2014). It has a small variation with a rate of 0.1 and −0.03 ppt/yr for China and the NH background, respectively. With some exemption applications, CH₂Br exhibited an increase of 0.3 and 0.1 ppt/yr for China and the NH background, respectively. Meanwhile, increase rates of HCFC-22 (8.3 ppt/yr), HCFC-141b (0.9 ppt/yr), HCFC-142b (0.1 ppt/yr) and HFC-134a (4.4 ppt/yr) indicate wide applications and notable emissions of these species in China. NH background of these species showed a similar trend but with lower values. For unregulated species, they did not show nationwide consistency because their lifetimes are relatively short and emissions vary significantly from region to region.

3.2. Air mass classification and its impacts on halocarbons

Backward trajectories were computed by the HYSPLIT model to understand regional transport of halocarbons for the summer and winter-spring periods. Mixing ratios of halocarbons associated with different air mass clusters were used to analyse the impacts of their source regions.

As shown in Fig. 6a, summertime air masses were grouped into four clusters, i.e. East China (EC) air mass, Northeast China (NEC) air mass, Siberia air mass and South China Sea (SCS) air mass. EC air mass occupied an overwhelming proportion of 60% due to the East Asia
summer monsoon. NEC, SCS and Siberia air masses accounted for 26%, 9% and 4%, respectively. SCS and Siberia air masses originated from thousand kilometres away with a faster speed, while EC and NEC air masses convoluted in small ranges and moved slowly. Fig. 6b shows, in the winter-spring period, Siberia, North China Plain (NCP) and Central Asia (CA) air masses split the trajectories with proportions of 53%, 34% and 12%, respectively. Due to Mongolia high pressure, Siberia and CA air masses came from the distant northwest, while the slowly-moving NCP air mass hovered over the polluted NCP region before reaching the site.

Median mixing ratios and standard deviations of halocarbons associated with each cluster are presented in Fig. 6c–d. Summertime Siberia air mass only contained 2 samples and was excluded from discussion. Compared with the NEC air mass, air originated from EC and SCS loaded higher levels of most halocarbon species. The coefficient of variation (CV) of CFCs, Halons and CH3CCl3 in these three air masses were between 0.005 and 0.05, indicating low measures of dispersion and suggesting they have distributed evenly in the atmosphere. Comparing the halocarbon levels in the EC, NEC and SCS air masses, concentrations of some industrial species (HCFCs, HFC-134a, C2HCl3, C2Cl4, C2H4Cl2) were on average 60% higher in the EC air mass. Meanwhile, levels of marine species (e.g., CH3I, CH2Br2, CHBrCl2, CHBr2Cl, CHBr3) were on average 52% higher in the SCS air mass. The vertical height variations of air masses provide information about the vertical mixing and pollutant accumulation processes. All air masses showed low speed and altitude before arriving at the sampling site. Such movement indicates weak turbulence diffusion and is favourable for accumulation of air pollutants. These air masses were much more
influenced by Jiangsu, Shandong and Hebei provinces. The EC air mass originated from the East China and slowly hovered above Jiangsu and Shandong provinces at altitudes below 1000 m before reaching the site. Thus, the EC air mass was well mixed with some industrial species when passing through these regions. The SCS air mass also passed through the south of Shandong Province with a low altitude below 1000 m, but it might not mix much with local pollutants due to the fast speed. Thus, the SCS air mass only carried some higher levels of marine species originating from the South China Sea. The NEC air mass slowly passed through the Bohai Sea and north of Shandong Province at altitudes below 2000 m before reaching the site. Compared with the southern area of Shandong Province, the Bohai Sea and the northern part of Shandong were relatively clean. So the NEC air mass carried the lowest levels of halocarbons of all three air masses.

In the winter-spring period, the NCP air mass dominated the pollutants of clusters. Concentrations of CFCs, Halons and CH₃CCl₃ were similar in different air masses, while mixing ratios of the rest species were on average 80% higher in the NCP air mass. The NCP air mass moved slowly in the polluted NCP region at altitudes below 2000 m, leading to strong accumulation of air pollutants in the boundary layer. This plume had strong regional influence and brought high levels of halocarbons to the YelRD region. In contrast, the Siberia and CA air

Fig. 5. Historical trends of selected halocarbons in China. Dashed lines represent curve fitting of annual variety for different regions. The NH background is the average of three sites (Trinidad Head, Mace Head and Junfraujoch). Data are the average values. But for the YelRD region, they are presented by the median to minimize the influences of extreme points. Historical data are from previous studies (An et al., 2012a; An et al., 2012b; Barletta et al., 2006; Chan and Chu, 2007; Chang et al., 2008; Fang et al., 2014; Fang et al., 2012a; Guo et al., 2009; Qin, 2007; Shao et al., 2011; Wang et al., 2014b; Wu et al., 2014; Wu et al., 2013; Yao et al., 2012a; Yao et al., 2012b; Zhang et al., 2010; Zhang et al., 2017; Zhang et al., 2014).
masses originated from Central Asia at an altitude of 2000 m and moved fast along the pathway. Thus, they carried the lowest levels of pollutants.

Overall, the backward trajectory cluster and air mass load show that the North China Plain and East China region were the main source regions of halocarbons that affected the YelRD in winter-spring and summer, respectively.

3.3. Source apportionment analysis

Fig. 7 shows the mean source profiles and contributions of each source derived from PMF. Five types of sources were finally identified according to goodness of fit to the data and prior knowledge.

The first source was characterized by high loadings of CH$_3$Cl and identified as biomass burning (Xue et al., 2011). About 84% of CH$_3$Cl was from this source. Biomass burning is also proposed as a minor source of CHCl$_3$ and CH$_2$Cl$_2$ (Sarkar et al., 2018). And indeed about 24% of CHCl$_3$ and 20% of CH$_2$Cl$_2$ originated from this source in this study.

The second source was distinguished as solvent usage. Its profile is mainly made up with high mass fractions of C$_2$HCl$_3$ (75%) and C$_2$Cl$_4$ (67%). C$_2$HCl$_3$ is primarily used as surface degreaser agent and solvent in electronic and textile industrial processes (Simmonds et al., 1998). C$_2$Cl$_4$ is widely used as industrial cleaning solvents in electronic industry (Guo et al., 2009). In addition, as shown in Fig. 3, the most polluted air masses of C$_2$Cl$_4$ and C$_2$HCl$_3$ were from the southwest (Dongying City) where some electronic and cleaning factories were located. C$_2$Cl$_4$ and C$_2$HCl$_3$ also showed a good correlation with each other ($r = 0.7$).

The third source was characterized as refrigerants and foam blowing...
by high mass percentages of HCFCs (HCFC-22, HCFC-141b and HCFC-142b), HFC-134a, and CFC-113. HCFC-22, HFC-134a and HCFC-142b are major refrigerant replacements in post-MP era (posterior to the signing of the Montreal Protocol). HCFC-22 is widely used in commercial refrigeration and refrigeration transport. HFC-134a is commonly applied in mobile air conditioning system (McCulloch et al., 2003). HCFC-142b is mainly used in air conditioning system in high-temperature environment and as a blowing agent (Simmonds et al., 1998). Meanwhile, HCFC-141b is primarily used as foam blowing agents in the manufacture of shoes and motor dashboards, and it also used as cleaning fluids for electrical equipment (Derwent et al., 2007). CFC-113 is mainly used as refrigerant and cleaning solvents in leather making and electronic industry (Fraser et al., 1996). There are many refrigeration plants and leather shoes factories in the YelRD region. In addition, with the rapid economic development, air conditioners and private vehicles become commonplace in this region, which are strong sources of HCFCs and HFC-134a species.

Leakage of CFCs banks were identified as the forth source for high proportions of CFC-11, CFC-12, CFC-113 and CCl4. These compounds have been phased-out for about 10 years in China (Aucott et al., 1999; McCulloch and Midgley, 2001; McCulloch et al., 2003). However, CFCs and CCl4 produced in the pre-MP era are stored in equipment and products and can leak into the environment.

The last source is marine air. This source is associated with high mass percentage of DMS, CHBr\textsubscript{3}, CHBr\textsubscript{2}Cl, CHBr\textsubscript{3}Cl and CH\textsubscript{2}Br\textsubscript{2}. These species are typically used as ocean tracers (Cox et al., 2003; Sarkar et al., 2018). They showed good correlation with each other (with $r$ ranging from 0.28 to 0.97), indicating they are from marine and marine biogenic sources.

Based on the PMF analysis, average contributions of each factor to the total halocarbons are illustrated in Fig. 8a. It can be seen that biomass burning (37%) and marine (23%) are two major contributors in this region. Refrigerants and foam blowing, solvent usage, and leakage of CFCs banks contributed 15%, 15% and 10%, respectively. Wheat straw burning can make significant contributions to halocarbons in summer as many farmlands exist in this region. The Bohai Sea is close to the YelRD region, and the influence of it on total halocarbons levels was also high. Compared with developed regions (the PRD, Singapore, South Korea), local industry of this area is relatively underdeveloped and showed much lower contributions (Guo et al., 2009; Li et al., 2014; Sarkar et al., 2018).

Considering the seasonal difference, PMF was also conducted separately for the summer and winter-spring periods. The results are shown in Fig. 8b and c. In the winter-spring period, five sources, namely, biofuel burning (39%), marine (32%), solvent usage (15%), refrigerants and foam blowing (8%), leakage of CFCs banks (6%) were identified. It is worth noting that biofuel burning was identified as the largest source in the profile with high percentage of CH\textsubscript{3}Cl (80%) and CO (49%). This source is the result of biofuel/biomass burning in northern China for residential heating in winter. While other sources had similar proportions to the above-identified overall source profiles. In summertime, biomass burning, refrigerants and foam blowing were major sources of halocarbons with proportions of 33% and 29%, respectively. Meanwhile, marine, solvent usage, and leakage of CFCs banks contributed 19%, 6% and 13%, respectively. The high proportions of biomass burning and refrigerants can be explained by wheat straw burning and air conditioning in summer. PMF results reveal that halocarbons in the YelRD region were mainly influenced by biomass (biofuel) burning, marine air and refrigerants.

4. Conclusions

A two-phase field campaign was conducted in the YelRD area which is China’s second largest estuary delta region. Halogenated hydrocarbons samples were collected at a rural site and in the oil fields in winter-spring and summer 2017. Concentrations of major MP eliminated species at the sampling site were comparable to the NH background, indicating the implementation of MP is effective in China. The high enhancement of CH\textsubscript{3}Br might be the result of the emissions from freshwater wetlands and coastal saltmarshes. Meanwhile, enhancements of MP under control species (HCFCs and HFC-134a) and unregulated species were moderate and significant. Oil production, transport and refining processes were found to have negligible influences on halocarbon levels at the site. Seasonal and diurnal varieties of
halocarbons in each category were different. The factors that led to those discrepancies were chemical lifetime, emission and meteorology differences as well as different phase-out schedules under the Montreal Protocol. Historical data show that in China, levels of regulated species exhibited a declining trend while concentrations of MP under control species showed an increasing trend. Back trajectory analysis indicates air masses from the North China Plain and East China loaded the highest concentrations of most halocarbons during the winter-spring and summer periods, respectively. PMF model results show that biofuel burning and biomass burning were the most major sources of total halocarbons in winter-spring and summer, respectively. Refrigerants and foam blowing, solvent usage, leakage of CFCs banks and marine air are also important sources of halocarbons in this region. This study demonstrates effective emission reductions of most MP eliminated species in China, and calls for further efforts to control the CFCs replacements and short-lived halocarbons.

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Appendix A. Supplementary data

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